

1 **Treatment of losses of ultrafine aerosol particles in long sampling tubes**  
2 **during ambient measurements**

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7 **Abstract**

8 Long sampling tubes are often required for particle measurements in street  
9 canyons. This may lead to significant losses of the number of ultrafine (those below  
10 100 nm) particles within the sampling tubes. Inappropriate treatment of these losses  
11 may significantly change the measured particle number distributions (PND), because  
12 most of the ambient particles, by number, exist in the ultrafine size range. Based on the  
13 Reynolds number ( $R_e$ ) in the sampling tubes, most studies treat the particle losses using  
14 the Gormley and Kennedy (1949) laminar flow model or the Wells and Chamberlain  
15 (1967) turbulent flow model. Our experiments used a particle spectrometer with  
16 various lengths (1.00, 5.47, 5.55, 8.90 and 13.40 m) of sampling tube to measure the  
17 PNDs in the 5–2738 nm range. Experiments were performed under different operating  
18 conditions to measure the particle losses through silicone rubber tubes of circular  
19 cross-section (7.85 mm internal diameter). Sources of particles included emissions  
20 from an idling diesel engine car in a street canyon, emissions from a burning candle  
21 and those from the generation of salt aerosols using a nebuliser in the laboratory.  
22 Results showed that losses for particles below  $\approx 20$  nm were important and were largest  
23 for the smallest size range (5–10 nm), but were modest for particles above  $\approx 20$  nm. In  
24 our experiments the laminar flow model did not reflect the observations for small  $R_e$ .  
25 This may be due to the sampling tubes not being kept straight or other complications.  
26 In-situ calibration or comparison appears to be required.

27 **Keywords:** Aerosol deposition, Particle diffusion losses; Particle number distribution;  
28 Street canyon; Ultrafine particles

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## 1    **1.    Introduction**

2            In many studies of ambient aerosols, experimental conditions force the use of  
3 long sampling tubes when commonly deployed instruments such as the Scanning  
4 Mobility Particle Sizers (SMPS), the Electrical Low Pressure Impactor (ELPI), the  
5 Aerodynamic Particle Sizer (APS) or the Differential Mobility Spectrometer (DMS500)  
6 are used for measuring particle number distributions (PND) and/or particle number  
7 concentrations (PNC). These studies include a recent experimental campaign (Kumar et  
8 al., 2008a–d) where a fast response differential particle spectrometer (DMS500)  
9 measured the PNDs at the roadside. Losses of ultrafine (those below 100 nm) particles  
10 inside the sampling tubes were observed, with smaller particles suffering greater losses.  
11 Correction for these losses (by number but not by mass) is critical since most of the  
12 ambient particles, by number, exist in the size range (below 100 nm) where these losses  
13 are also the largest (Kumar et al., 2008b). Based on the Reynolds number in the  
14 sampling line, several studies (Agus et al., 2007; Kumar et al., 2008a; Kumar et al.,  
15 2008c; Lingard et al., 2006) treat particle losses in straight tubes using either the  
16 Gormley and Kennedy (1949) laminar model or the Wells and Chamberlain’s (1967)  
17 turbulent model. Unfortunately, only a few studies discuss the importance of such  
18 losses (Noble et al., 2005; Symonds et al., 2007a; Wang et al., 2002). Noble et al.  
19 (2005) measured particles in a continuous field measurement for laminar flow  
20 conditions in sampling tubes using a SMPS and an APS through a 3 m long aluminium  
21 tube (2.54 cm internal diameter) that was vertically oriented with slight bends for  
22 connections to the sampling instruments. They reported that the penetration efficiency  
23 was substantially smaller than the theoretical penetration efficiency for both ultrafine  
24 and fine particles. Similarly, Symonds et al. (2007a) reported that the penetration  
25 efficiency of particles in the 5–100 nm range for sampling tubes up to 25 m long, made  
26 of silicone rubber, was much closer to the turbulent flow model even though the

1 Reynolds number indicated laminar flow. Wang et al. (2002) reported the effect of  
2 bends and elbows on diffusion of mono disperse particles (5–15 nm) for a range of  
3 Reynolds numbers (80–950). They concluded that in a flow passage with four elbows,  
4 each having 90° bends, the penetration efficiency was up to 44% smaller than for a  
5 straight tube of same length. However, particle losses and their appropriate treatment  
6 for different length of sampling tubes, as would typically be used in ambient aerosol  
7 studies, need further attention.

8 In this study, a DMS500, with a four-way switching system, measured the  
9 PNDs in a broad (5–2738 nm) size range, pseudo-simultaneously. Various lengths of  
10 sampling tubes, between 1 and 14 m, made of silicone rubber, having 7.85 mm internal  
11 diameter (i.d.), were used to measure the particle penetration efficiencies. Three  
12 different and continuous sources of particles were used: an idling diesel-engined car in  
13 a street canyon, a burning candle and salt aerosols from a nebuliser (PARI LC+) in the  
14 laboratory.

15 The aim of the study was to investigate particles losses in sampling tubes of  
16 various lengths under different operating conditions of the DMS500. The experimental  
17 results are compared with particle loss models for laminar and turbulent flow and it is  
18 shown that such losses are extremely important for the ultrafine particles that are  
19 dominant (by number) in the urban environment (Kumar et al., 2007, 2008a–d; Longley  
20 et al., 2003).

## 21 **2. Methodology**

### 22 **2.1 Theories of particle losses in sampling line**

23 There are five main mechanisms which may lead to particle losses on to the  
24 surface of a sampling tube; these are sedimentation (gravitational), thermophoresis,  
25 electrostatic, inertial impaction and diffusion (Friedlander, 2000; Hinds, 1999; Ketznel

1 and Berkowicz, 2004). Of all potential losses, those due to diffusion and inertial  
 2 impaction are most important for ambient particle measurements (Hinds, 1999). The  
 3 second of these is only important under turbulent flow conditions and for particles  
 4 larger than 100 nm (Lee and Gieseke, 1994). Gormley and Kennedy (1949) first  
 5 derived the equation for diffusional losses in a fully developed laminar flow (Reynolds  
 6 number,  $Re < 2300$ ) through a tube of circular cross section (diameter  $d_t$ ) with a uniform  
 7 inlet PNC, as a function of a dimensionless deposition parameter  $\beta = 4DL/\pi d_t^2 \bar{U}$ ,  
 8 where  $D$  is the diffusion coefficient (see supplementary Section S.1 for details),  $L$  is the  
 9 tube length and  $\bar{U}$  is the average flow velocity through the tube. A simplified version  
 10 of a more complicated and more accurate expression (Hinds, 1999), gives penetration  
 11 efficiency  $P$ , which is the fraction of entering particles ( $N_{in}$ ) that exit ( $N_{out}$ ) through a  
 12 tube, with an accuracy of 1% for all values of  $\beta$  to be:

$$13 \quad P = N_{out}/N_{in} = 0.819e^{-11.5\beta} + 0.0975e^{-70.1\beta} \quad \text{for } \beta \geq 0.009 \quad (1a)$$

$$14 \quad P = N_{out}/N_{in} = 1 - 5.50\beta^{2/3} + 3.77\beta \quad \text{for } \beta < 0.009 \quad (1b)$$

15 For straight sampling tubes, Ramamurthi et al. (1990) confirmed the accuracy  
 16 of the above equations during their study that used radioactive  $^{218}\text{PoO}_x$  aerosol clusters  
 17 through a 2.2 cm i.d. tube of various lengths (i.e., 88 mm, 205 mm and 317 mm). Their  
 18 work was supported by Alonso et al. (1997) for particles down to 2 nm diameter by  
 19 analysis of the penetration efficiencies of nanometre-sized aerosol particles through a  
 20 plastic tube of 10 mm i.d.

21 Under turbulent pipe flow conditions ( $Re > 2300$ ), the deposition onto tube  
 22 surfaces is more complicated and equations describing it cannot be solved explicitly.  
 23 Assuming that turbulent flow provides a constant concentration everywhere beyond a  
 24 thin boundary layer next to the surface of the tube walls where the flow is laminar and  
 25 the concentration decreases linearly from constant to zero at the surface walls, Wells

1 and Chamberlain (1967) gave an expression for diffusive deposition velocity ( $V_d$ )  
2 through the laminar sub-layer for turbulent flow in a tube as:

$$3 \quad V_d = \frac{0.04\bar{U}}{R_e^{1/4}} \left( \frac{\rho_g D}{\eta} \right)^{2/3} \quad (2)$$

4 where  $\rho_g$  and  $\eta$  are the density and the viscosity of fluid passing through the tube.

5 Using the deposition velocity (as given in Eq. 2), Lee and Gieseke (1994) presented an  
6 empirical equation for the overall penetration through a tube of length  $L$  subjected to  
7 losses to the walls by diffusion and inertial impaction from turbulent flow as:

$$8 \quad P = \frac{N_{out}}{N_{in}} = \exp\left(\frac{-4V_d L}{d_t \bar{U}}\right) \quad (3)$$

9 Eq. (3) also includes inertial deposition but this is only significant for particles larger  
10 than around 100 nm.

## 11 **2.2 Instrumentation**

12 A four-way solenoid switching system was used with a fast response  
13 differential mobility spectrometer (Cambustion DMS500) to measure the particle  
14 number and size distributions in 5–2738 nm size range through different lengths of tube  
15 pseudo-simultaneously. Cylindrical sampling tubes of 7.85 mm i.d. and various  
16 lengths,  $L_{base}$  (1.00 m),  $L_1$  (5.47 m),  $L_2$  (5.55 m),  $L_3$  (8.90 m) and  $L_4$  (13.40 m), were  
17 used to analyse the particle losses inside the electrically conductive sampling tubes that  
18 were made of silicone rubber. As seen in Fig. 1, there were two parts to the switching  
19 system, namely a stainless steel manifold having one inlet and four outlets, and a four-  
20 way solenoid switching system. A steel funnel was fixed at the head of steel manifold  
21 inlet. The outlets of the manifold were connected to cyclones by small pieces of rubber  
22 sampling tubes. The cyclones were then connected to the four-way switching system  
23 by another piece of silicone rubber sampling tube; the length of these small connecting

1 sections is included in above-mentioned tube lengths  $L_{\text{base}}$ ,  $L_1$ ,  $L_2$ ,  $L_3$  and  $L_4$ . Finally,  
2 the four-way switching system was connected to the DMS500, as seen in Fig. 1. The  
3 cyclones prevent large particles entering the sampling tube and the instrument. All the  
4 sampling tubes were laid horizontally on the ground but small bends were present due  
5 to the difference in height between the instrument and the emission sources of particles.

6 The DMS500 used is capable of measurements over two size ranges (i.e., 5–  
7 1000 nm and 5–2738 nm). Each of these ranges requires different set-points for the  
8 instrument's internal flows, voltages and pressure. The use of these two ranges in these  
9 experiments enabled two different sample flow rates and pressures to be examined  
10 (discussed in detail in Section 3.2). To measure the PNDs in the 5–1000 nm range, steel  
11 restrictors with holes of  $\approx 1.00$  mm i.d. were placed upstream of the cyclones to  
12 maintain a flow rate of 8 standard litres/min (slpm) and a pressure of 0.25 bar (the same  
13 as the instrument's classification column) inside the sampling tube. However, the  
14 instrument could also be used to measure PNDs in the 5–2738 nm range, when  
15 restrictors of 0.5 mm i.d. were substituted for the 1.00 mm ones, maintaining a flow  
16 rate of 2.5 slpm and pressure 0.16 bar inside the sampling tube (the classifier operates  
17 at lower pressure to achieve the increased electrical mobility size range). The orifice  
18 plates (a) reduce the pressure inside the sampling tubes hence reducing the likelihood  
19 of particle agglomeration, (b) set the sample flow rate to that required by the  
20 instrument, and (c) improve the time response of the instrument.

## 21 **2.2 Measurements**

22 Three different sources for nearly steady-state emissions were selected to  
23 measure the particle losses through the different lengths of sampling tubes. Exhaust  
24 emissions from a stationary diesel engine car at idle (Model: Rover 25 TD) in a street  
25 canyon (500 mm from the tail pipe) mimicked real field (i.e., operational) conditions.  
26 To verify these results, further experiments were replicated in the laboratory (i.e.,

1 controlled) conditions using particle emissions from a burning tea light candle and salt  
2 aerosols generated by a nebuliser (Pari LC+) at two different pressures (0.5 and 2.0 bar)  
3 to change the aerosol concentration. The details of the use of this nebuliser to generate  
4 salt aerosols have been described elsewhere (Fennell et al., 2007). The tea light candle  
5 consisted circular plugs of paraffin wax (37 mm diameter and 15 mm high) in a thick  
6 aluminium casing (Daeid and Thain, 2002). The DMS500 measures particles based on  
7 electrical mobility equivalent diameter (hereafter called as  $D_p$ ) that implicitly takes into  
8 account the characteristics of aerosols (i.e., shape, surface area, charge etc.). The  
9 different types of particle sources were found not to influence the particle losses in this  
10 study, as is discussed later in Section 3.2 (Figs. 3–5) where experimental data from all  
11 sources show similar penetration efficiency for a particular particle diameter. This is  
12 because the equations governing losses of particles (see Section S.1 in supplementary  
13 material) are based on the aerodynamic mobility diameter, which is very close to  $D_p$ , so  
14 that differences between different aerosols are explicit in the models. Measurements  
15 were taken at a sampling frequency of 0.5 Hz, rather than the maximal frequency of 10  
16 Hz to improve the signal-to-noise ratio, continuously for 1-min in each tube length.  
17 These were repeated for four complete cycles (i.e. four  $\times$  1-min measurements for each  
18 tube length taken on four different occasions). Further details of the experiment and the  
19 operating conditions are shown in Table 1.

## 20 **3. Results and Discussion**

### 21 **3.1 Effect of tube length on particle number distributions**

22 To derive the losses in the sampling tubes, it is assumed that the losses in  $L_{base}$   
23 are equivalent to the losses in the first metre of each of other four tubes, and subsequent  
24 losses are determined in each tube relative to their “corrected” length (i.e., their actual  
25 length minus 1 m). The size dependent penetration efficiencies for the effective length

1 of different tubes were then defined as the number concentration through the effective  
2 lengths  $L_1$ ,  $L_2$ ,  $L_3$  and  $L_4$  of tube divided by the number concentration penetrating  $L_{\text{base}}$ .  
3 Fig. 2 shows the effect of various length of sampling tubes on particulate emissions  
4 from a diesel–engined car. As the tube lengths increases from  $L_{\text{base}}$  to  $L_4$ , the magnitude  
5 of the PNDs decreased. The PNDs in smaller size range (below  $\approx 20$  nm) showed larger  
6 changes, indicating relatively greater losses of particles with increased length of  
7 sampling tubes. However, the PNDs above  $\approx 20$  nm showed negligible changes,  
8 indicating modest particle loss. The PNCs, which were obtained by integrating the  
9 PND profiles seen in Fig. 2, decreased about 3, 7, 28 and 32% for  $L_1$ ,  $L_2$ ,  $L_3$  and  $L_4$ ,  
10 respectively, with reference to  $L_{\text{base}}$ . It should be noted that the majority of the decrease  
11 in PNC was for particles below  $\approx 20$  nm where the maximum losses are also expected  
12 due to the higher diffusivity of smaller particles. For example, penetration efficiencies  
13 for  $L_4$  were observed to be  $\approx 10\%$  for 5 nm diameter,  $\approx 40\%$  for 10 nm diameter, and  
14 well above 60% for particles in the 10–20 nm size range. Other sources also showed  
15 similar trend for penetration efficiencies, as can be seen from Figs. 3–5.

### 16 **3.2 Effect of operating conditions on the size–dependent penetration of** 17 **particles in various tube lengths**

18 Size–dependent penetration efficiencies of particles in  $L_2$ ,  $L_3$  and  $L_4$  for all cases  
19 described in Table 1 are shown in Figs. 3, 4 and 5, respectively. The penetration  
20 efficiencies shown in these figures are estimated based on the PNCs which were  
21 averaged over the time period of four complete cycles as mentioned in Section 2.2.  
22 Particle emissions from various sources were nearly in steady–state conditions since the  
23 geometric mean diameter with standard deviation over the averaged periods were  
24  $13.7\pm 0.7$ ,  $12.5\pm 0.5$  and  $10.2\pm 0.8$  nm for car emissions, salt aerosols and candle  
25 emissions, respectively. Tube length  $L_1$  is not considered for further analysis since  $L_1 \approx$



1  $L_2$ . The results for  $Car_2$  and  $Car_3$  are generally similar with the penetration efficiencies  
2 being similar to the theoretical “turbulent” flow model except for  $Car_3$  with the longest  
3 sampling tube and for the larger particles. The candle results for all three sampling tube  
4 lengths also produce a similar result to the  $Car_2$  and  $Car_3$  result.

5 Sensitivity analysis of Eqs. (1–3) suggest that apart from the length of the  
6 sampling tubes, their diameter, sample flow rate or mean velocity (since both are  
7 linked), sample line pressure and temperature are variables affecting the particle losses  
8 in the sample line. The effect of different tube diameters on penetration efficiencies  
9 could not be seen because our experiments used only a fixed diameter tube. The effect  
10 of sample line temperature (assumed to be equal to ambient temperature) on  
11 penetration efficiencies seem to be modest for our conditions. This can be seen by the  
12 comparison of Figs. 3a, 4a and 5a (for 8.2 °C) with Figs. 3b, 4b and 5c (for 26 °C),  
13 respectively, where all other conditions were similar except temperature (see Table 1).  
14 For instance, the average of the modelled penetration efficiency in the 5–100 nm size  
15 range for  $L_2$  showed fractional changes;  $\approx 89\%$  for 8.2 °C as compared with  $\approx 88.8\%$  for  
16 26 °C, as seen in Figs. 3a and 3b. However, this difference was larger for experimental  
17 results where the average penetration efficiencies were  $\approx 78\%$  for 8.2 °C as compared  
18 with  $\approx 73\%$  for 26 °C.

19 When using the DMS500, the sample line pressure and the sample flow rate are  
20 not independent variables, due to the requirement for 8 slpm when 0.25 bars is used, or  
21 2.5 slpm with 0.16 bars. Considering our experimental conditions (as explained in  
22 Table 1 for different cases) the sample flow rate was identical (2.5 slpm) for the first  
23 four cases, and these cases are averaged and shown in part (d) of each of Figs. 3–5  
24 whereas this was 8 slpm for the last two cases, as seen in Table 1 and in part (c) of  
25 Figs. 3–5. Comparison of Figs. 3c, 4c and 5c with Figs. 3d, 4d and 5d, respectively,

1 indicate the effect of change in the sample flow rate on the measured and modelled  
2 results. As expected from the sensitivity analysis, the penetration efficiencies for  
3 modelled results for all lengths of sampling tubes were consistently larger for 8 slpm  
4 sample flow rate (Figs. 3c, 4c and 5c) than for 2.5 slpm sample flow rate (Figs. 3d, 4d  
5 and 5d) since the diffusive deposition velocity, as seen in Eq. (2), increases  
6 considerably with increased sample flow rate. For example, the average of the  
7 modelled penetration efficiency in the 5–100 nm size range for  $L_2$  during laminar flow  
8 conditions was  $\approx 93\%$  for 8 slpm as compared with  $\approx 88\%$  for 2.5 slpm, as seen in Figs.  
9 3c and 3d. Surprisingly, this effect was not clearly distinguishable on the measured  
10 penetration efficiencies which were much smaller (and nearly similar for both sample  
11 flow rates) than those for modelled penetration efficiencies for laminar flow conditions;  
12  $\approx 74\%$  for 8 slpm as compared with  $\approx 73\%$  for 2.5 slpm. Moreover, the average of the  
13 measured penetration efficiencies in the 5–100 nm size range was found to be closer to  
14 those for the turbulent flow model, where these were  $\approx 60\%$  for 8 slpm than  $\approx 54\%$  for  
15 2.5 slpm. The reasons for this are discussed in Section 3.3. Similar observations were  
16 found for other lengths ( $L_3$  and  $L_4$ ) of the sampling tubes, as can be seen from Figs. 4–  
17 5.

18 Further investigation of the measured results showed that particle losses are  
19 very significant between 5 and 10 nm for all cases (Figs. 3–5); these are as high as  
20  $\approx 90\%$  for 5 nm particles and  $\approx 60\%$  for 10 nm diameter. There is a sharp rise in  
21 penetration efficiency between 10 nm and 20 nm diameter. Above 20 nm diameter  
22 particle losses are modest.

### 23 **3.3 Comparison of experimental results with laminar and turbulent flow** 24 **diffusion models**

25 Particle losses for various cases are compared below with diffusion models for

1 laminar (Hinds, 1999) and turbulent (Wells and Chamberlain, 1967) flow, which are  
2 represented by Eqs. (1) and (3), respectively. As shown in Table 1, flow conditions in  
3 the sampling tubes were laminar ( $R_e = 461$  and  $1409$ ) during all the experiments. For all  
4 cases, particle losses were substantially larger for the experimental results than for  
5 modelled results calculated for laminar flow, as seen in Figs. 3–5. Interestingly, all the  
6 cases presented in Figs. 3, 4 and 5 for  $L_2$ ,  $L_3$  and  $L_4$ , respectively, show that measured  
7 penetration efficiencies for particles below  $\approx 20$  nm diameters are generally closer to the  
8 modelled diffusion losses for turbulent flow. Conversely, the measured penetration  
9 efficiencies for particles above  $\approx 20$  nm indicated modest particle losses, but part (d) of  
10 Figs. 3–5, interestingly, shows that the measured penetration efficiencies are closer to  
11 the modelled penetration efficiencies for the laminar flow; no clear explanation for this  
12 similarity was found. Particle losses below  $\approx 20$  nm only are discussed further in this  
13 article.

14 As discussed in Section 2.1, studies for short and straight tubes (Alonso et al.,  
15 1997; Ramamurthi et al., 1990) confirmed the laminar flow model under conditions  
16 commonly found *within* aerosol instruments. Conversely, when relatively longer  
17 sampling tubes are used, particle losses where laminar flow is expected have been  
18 found to be considerably larger than the modelled losses for laminar flow conditions  
19 (Wang et al., 2002; Noble et al., 2005; Symonds et al., 2007), as is confirmed by our  
20 brief study. The present study tested these findings for various operating conditions and  
21 for different lengths of sampling tubes ranging from 1 to 14 m, as described in Table 1.  
22 It might be expected that losses greater than those which are predicted by the model  
23 could result from other mechanisms such as in the connections to the switching system.  
24 However, these can be neglected here since they are common to all sampling tubes.

25 Prediction of particle penetration efficiencies is not straightforward since they  
26 depend on the complex flow field inside the tubes. It is probable that non-straightness

1 (or small bends) in the sampling tubes will produce secondary flows, and intermittent  
2 enhanced movement of the particles towards the sampling tube walls. These flow  
3 conditions can also occur due to any roughness in the inner-walls of the sampling line;  
4 however, this effect will be relatively small and can be ignored (Pope, 2003). The  
5 evidence of the presence of secondary flow can be supported by arguments based on  
6 the Dean number (De). The strength of secondary flow produced by flow through a  
7 smooth bend of radius  $R$  can be represented by  $De = Re (d_t/2R)^{0.5}$ ; where  $d_t$  is the tube  
8 diameter (Pui et al., 1987). An increase in De will increase the strength of the  
9 secondary flow (Tsai and Pui, 1990), suggesting that for a fixed tube diameter an  
10 increase in angle of bend decreases the bend radius and increases De. For example  
11 increasing the angle of bend from  $45^\circ$  to  $90^\circ$  reduces the radius of bend by  $\approx 31\%$  and  
12 increases De by  $\approx 13\%$ . This suggests that an increase in angle of bend could result in  
13 much smaller particle penetration efficiencies than expected during laminar flow  
14 conditions. Furthermore, change in diameter of the sampling tubes is equally important  
15 for changing the De. For example, an increase in tube diameter from 7.85 mm to 10  
16 mm ( $\approx 22\%$ ) decreases  $Re$  by  $\approx 22\%$ , resulting in a decrease of  $\approx 11\%$  in De; however our  
17 experiments were limited to sampling tubes of fixed diameter.

18         The presence of secondary flow in sampling tubes is further substantiated by the  
19 fact that the measured penetration efficiencies are closer to the modelled penetration  
20 efficiencies from Wells and Chamberlain's (1967) expression (Eq. 3) that takes in to  
21 account the diffusion of particles under turbulent flow conditions.

22         The above arguments suggest that an individual lay-out and diameter of a  
23 longer sampling tube are important aspects for changing the flow conditions.  
24 Consequently, in-situ calibration of the particle penetration is likely to be essential for  
25 the appropriate treatment of particle losses, though our results show that this can be

1 conducted in the laboratory, prior to any field campaign, since the size dependent  
2 penetration is similar for all aerosols investigated, in accordance with theory.

#### 3 **4. Summary and conclusions**

4 Experiments were made under different operating conditions of the DMS500  
5 particle spectrometer. Irrespective of any measured size range and different operating  
6 conditions used in various set of experiments, it was found that particle losses for  
7 particles smaller than  $\approx 20$  nm were important and needed to be taken into account  
8 when using long sampling tubes ( $>1$  m). Maximal particle losses were found for  
9 particles between 5 and 10 nm, whereas losses were modest for particles larger than  
10  $\approx 20$  nm. It can be concluded that ignoring these losses, especially below 20 nm (where  
11 a substantial number of the particles in ambient aerosols are to be found) may  
12 appreciably change the measured PNDs in the ambient environment. It was also  
13 apparent that even when the Reynolds number indicated that the flow was laminar, the  
14 turbulent penetration model of Hinds described particle losses best. Of course, it is  
15 most prudent to determine the losses of particles for any particular experimental setup  
16 directly.

17 This study presents preliminary results, and it would be interesting to perform  
18 similar experiments for different diameter of tubes and for various Reynolds numbers.  
19 More importantly, a study producing a thumb rule for angle of bends at which particle  
20 losses are minimum and/or are close to laminar flow model, will greatly benefit  
21 ambient aerosol studies.

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1 **Figure Captions**

2 **Fig. 1.** Schematic diagram of experimental set up.  $L_{\text{base}}/L_1$ ,  $L_2$ ,  $L_3$  and  $L_4$  are total length  
3 of sampling tubes between the manifold and DMS500; the tube length between the  
4 four-way switching system and the DMS500 was 0.30 m.

5 **Fig.2.** Effect of various lengths of sampling tubes on a typical diesel car exhaust  
6 emissions; figure represents cases Car<sub>1</sub> and Car<sub>2</sub>.

7 **Fig. 3.** Size dependent measured and modelled penetration in **sampling tube  $L_2$**  for (a)  
8 Car<sub>2</sub> and Car<sub>3</sub> ( $P = 160$  mb,  $T_a = 8.2$  °C,  $Q = 2.5$  slpm,  $R_e = 461$ ), (b) Cand and Salt<sub>1</sub> ( $P$   
9  $= 160$  mb,  $T_a = 26$  °C,  $Q = 2.5$  slpm,  $R_e = 461$ ), (c) Salt<sub>2</sub> and Salt<sub>3</sub> ( $P = 250$  mb,  $T_a = 26$   
10 °C,  $Q = 8$  slpm,  $R_e = 1409$ ), and (d) average of (a) and (b). Acronyms  $P$ ,  $T_a$ ,  $Q$  and  $R_e$   
11 represent the sample line pressure, ambient temperature, sample flow rate and Reynolds  
12 number, respectively. Further descriptions of all cases plotted in Figs. a–d are described  
13 in Table 1.

14 **Fig. 4.** Size dependent measured and modelled penetration in **sampling tube  $L_3$**  for (a)  
15 Car<sub>2</sub> and Car<sub>3</sub> ( $P = 160$  mb,  $T_a = 8.2$  °C,  $Q = 2.5$  slpm,  $R_e = 461$ ), (b) Cand and Salt<sub>1</sub> ( $P$   
16  $= 160$  mb,  $T_a = 26$  °C,  $Q = 2.5$  slpm,  $R_e = 461$ ), (c) Salt<sub>2</sub> and Salt<sub>3</sub> ( $P = 250$  mb,  $T_a = 26$   
17 °C,  $Q = 8$  slpm,  $R_e = 1409$ ), and (d) average of (a) and (b). Further descriptions of all  
18 cases plotted in Figs. a–d are described in Table 1.

19 **Fig. 5.** Size dependent measured and modelled penetration in **sampling tube  $L_4$**  for (a)  
20 Car<sub>2</sub> and Car<sub>3</sub> ( $P = 160$  mb,  $T_a = 8.2$  °C,  $Q = 2.5$  slpm,  $R_e = 461$ ), (b) Cand and Salt<sub>1</sub> ( $P$   
21  $= 160$  mb,  $T_a = 26$  °C,  $Q = 2.5$  slpm,  $R_e = 461$ ), (c) Salt<sub>2</sub> and Salt<sub>3</sub> ( $P = 250$  mb,  $T_a = 26$   
22 °C,  $Q = 8$  slpm,  $R_e = 1409$ ), and (d) average of (a) and (b). Further descriptions of all  
23 cases plotted in Figs. a–d are described in Table 1.



## 1 Tables

2 **Table 1.** Summary of experimental and operating conditions.

Cases	Aerosol Source	Place of Sampling	Length of Sampling tubes used	Sampling conditions
Car <sub>1</sub>	Diesel car	50 cm away from exhaust tail pipe	$L_{base}$	Particle size range 5–2738 nm, Sample line pressure 160 mb Ambient Temperature 8.2 °C Sample flow rate 2.5 slpm Reynolds number ( $R_e$ ) = 461
Car <sub>2</sub>	Diesel car	50 cm away from exhaust tail pipe	$L_1, L_2, L_3$ and $L_4$	
Car <sub>3</sub>	Diesel car	At the exit (i.e., 10 cm away from exhaust tail pipe)	$L_{base}, L_2, L_3$ and $L_4$	
Cand	Candle	Candle flame 15 cm below the manifold inlet	$L_{base}, L_2, L_3$ and $L_4$	Particle size range 5–2738 nm Sample line pressure 160 mb Indoor Temperature 26 °C Sample flow rate 2.5 slpm $R_e = 461$
Salt <sub>1</sub>	Nebulised NaCl(aq)	Generated at 2 bar pressure	$L_{base}, L_2, L_3$ and $L_4$	
Salt <sub>2</sub>	Nebulised NaCl(aq)	Generated at 2 bar pressure	$L_{base}, L_2, L_3$ and $L_4$	Particle size range 5–1000 nm Sample line pressure 250 mb Indoor Temperature 26 °C Sample flow rate 8 slpm $R_e = 1409$
Salt <sub>3</sub>	Nebulised NaCl(aq)	Generated at 0.5 bar pressure	$L_{base}, L_2, L_3$ and $L_4$	

3

Figure 1.ppt

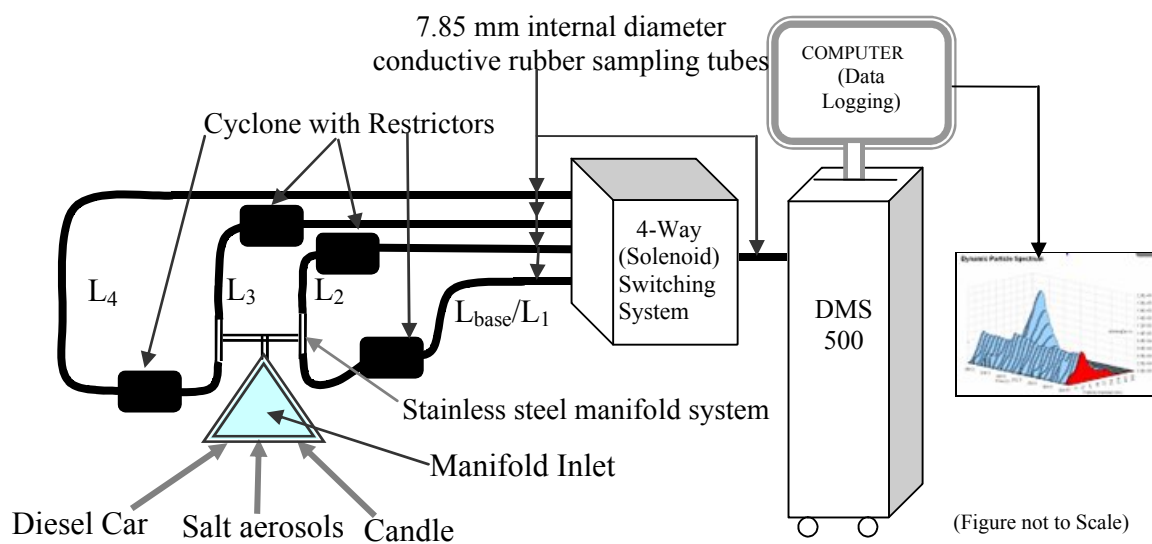


Figure 2.ppt

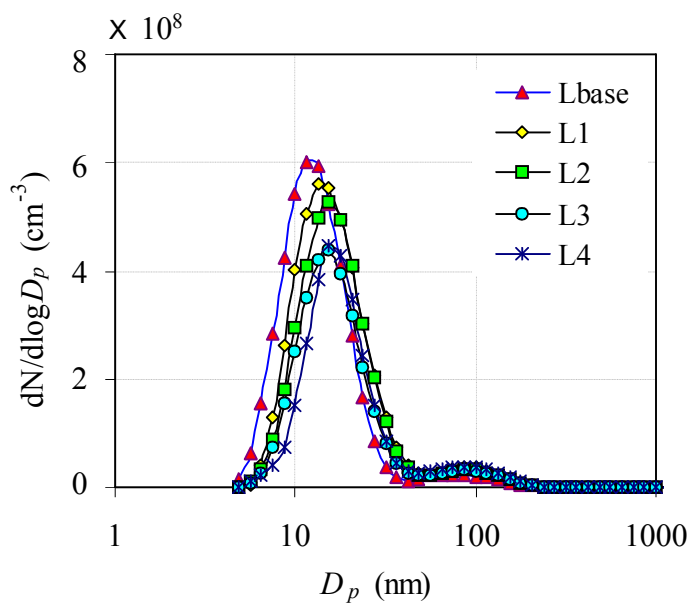


Figure 3\_For L2.ppt

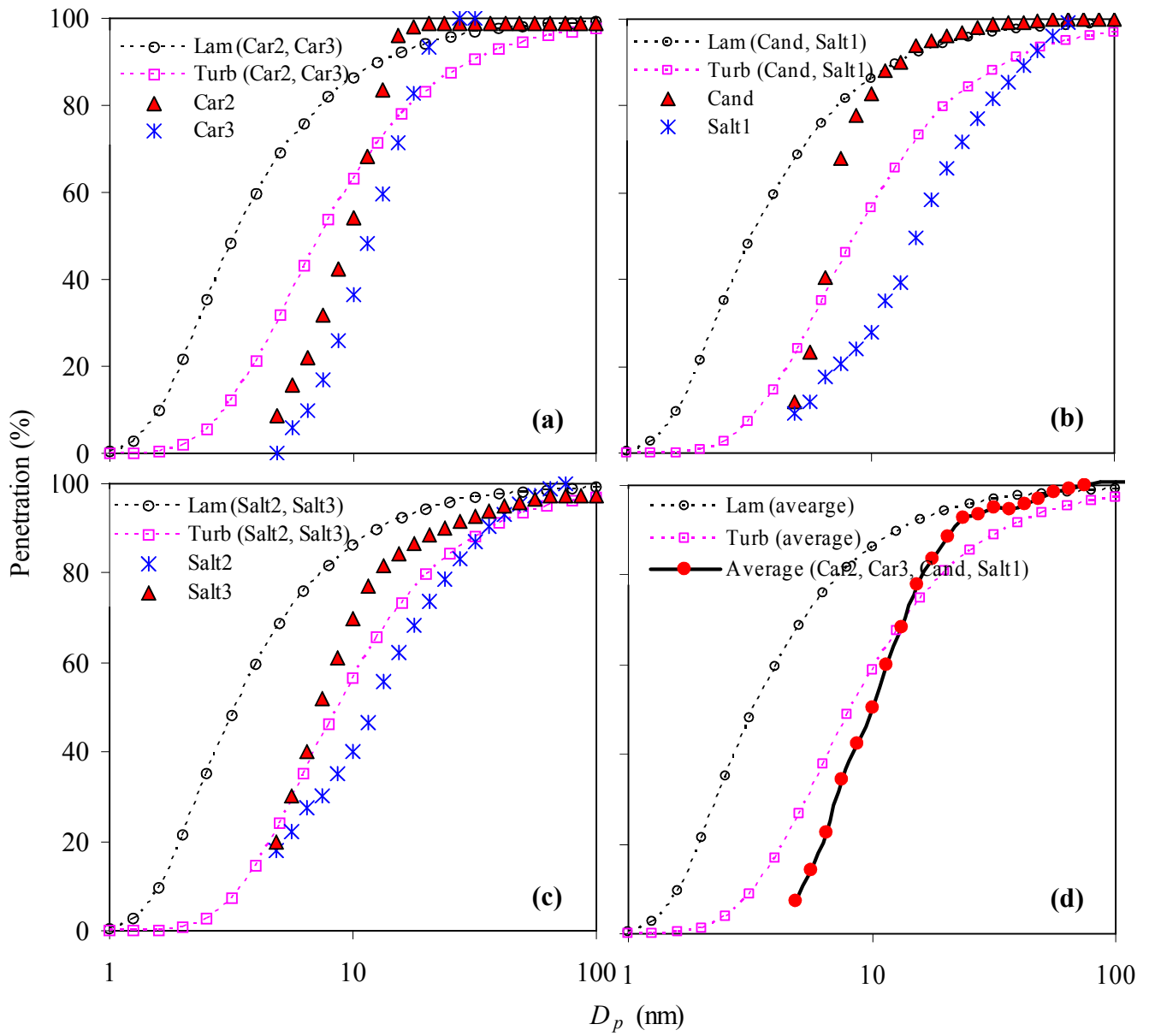


Figure 4\_For L3.ppt

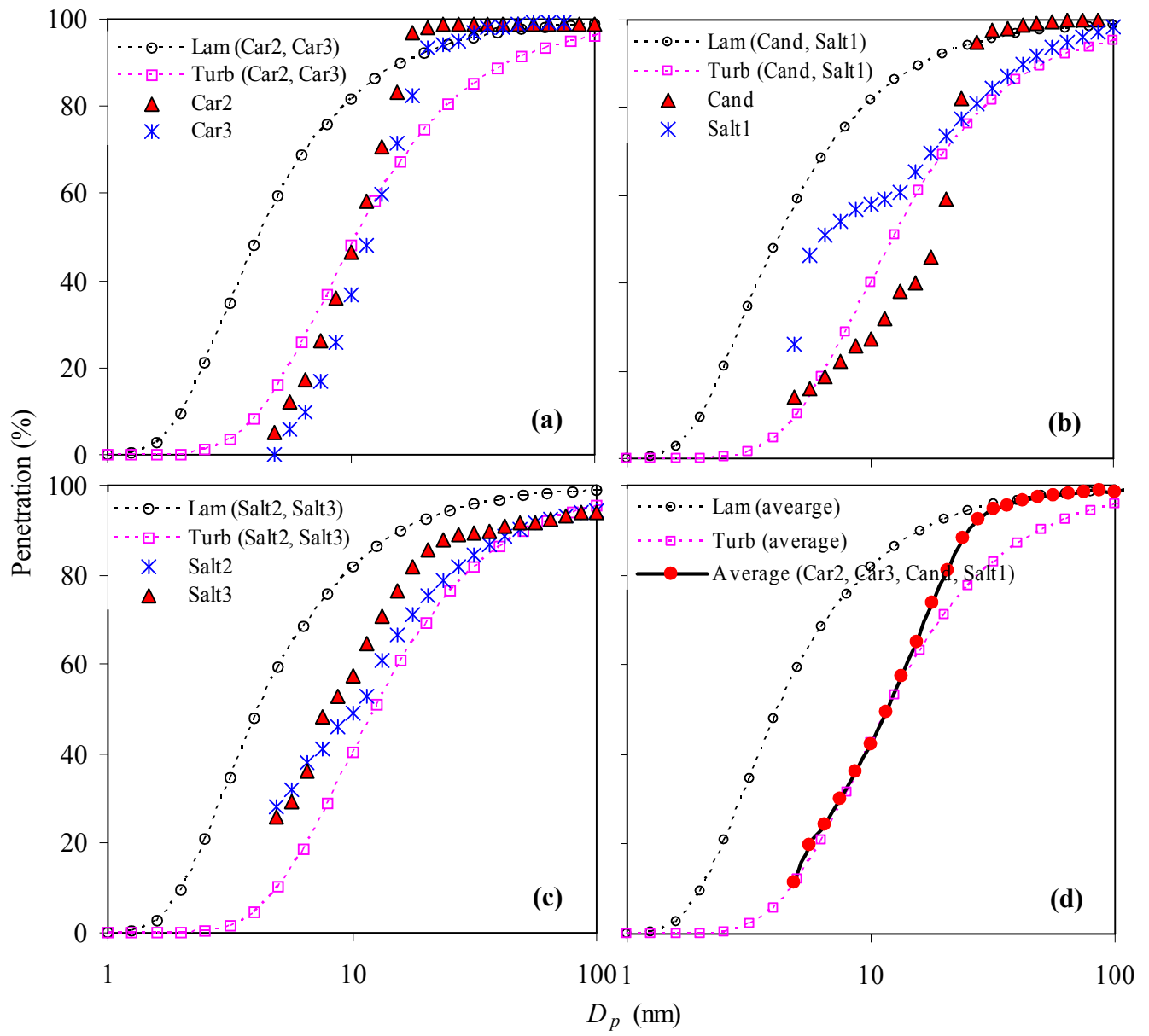


Figure 5\_For L4.ppt

